

Perspective

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Many pasts, many futures: Navigating the complexities of species reshuffling to help prevent extinctions

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Abstract

Preventing human-caused extinctions is a foundational aim of conservation. However, in addition to causing extinctions, humans have moved numerous species to new areas. A considerable percentage of these are threatened in their native ranges. Broadening our conservation ethos to include introduced species is contentious and requires critical thinking in empirical and normative dimensions to negotiate between conflicting conservation goals. Here, we present a series of questions to inspire critical thinking in the negotiation of these conflicts. Empirically, we suggest that conservationists should consider whether the effects of introduced species are due to their non-nativeness *per se* or are simply a consequence of the organism having a metabolism and taking up space. Importantly, this requires proper scientific comparison to the effects of similar native organisms – otherwise many claims of ‘harm’ are unfalsifiable and could be used to justify the eradication of any organism. We further propose questions to help conservationists sort facts from normative values, which often wear empirical clothes. Through empirical rigor, value transparency and critical justification of these values, we believe that twenty-first century conservation can become a future-facing and pluralistic discipline with a heightened ability to prevent extinctions in an increasingly unpredictable and novel biosphere.

Impact statement

The world is rapidly changing due to land use intensification, climate change and ongoing globalization. These changes are driving the emergence of novel ecosystems composed of native and introduced organisms, which are likely to expand as species migrate in response to changing environmental conditions. Many of these introduced organisms are threatened or extinct in their native ranges, and all of them are evolving. The possibility that these systems and these organisms may have conservation value is highly contentious for both empirical and normative reasons. We here present a series of questions to help guide critical thinking in both empirical and normative domains. This perspective aims to foster good-faith discussion around what we consider to be among the most salient challenges facing conservation in our modern world.

Introduction

For hundreds of thousands of years, camels (*Camelus spp.*) roamed Eurasia and northern Africa alongside straight-tusked elephants, *Stephanorhinus* rhinos, equids and large bovids. In North America, two to three species of wild equid, including modern horses (*Equus ferus*), grazed alongside mammoths and ground sloths. Rich megafauna communities were the norm on all continents for 50–30 million years until 50,000–12,000 years ago. The extinction of these large animals as humans spread from Africa appears to be one of the earliest human effects on the global biosphere (Svenning et al., 2024).

However, relationships between humans and nonhumans are complex. Some of the populations that went ‘extinct’ were later spread around the world in domesticated form, where they subsequently went ‘feral’ (Lundgren et al., 2018) – a form of spontaneous rewilding. From this, feral horses (*E. ferus*) have returned to North America, and the world’s only wild population of dromedary camels (*Camelus dromedarius*) roam in central Australia, a landscape recently populated by hippo-sized migratory diprotodonts, giant wombats and horse-like short-faced kangaroos (Prideaux et al., 2009; Price et al., 2017; Faurby et al., 2018). If feral camels were to be

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eradicated from Australia, where many biologists consider them to be ‘invasive’ species, we would make another species extinct in the wild.

Overall, 50% of introduced megafauna (>100 kg body mass) are threatened or extinct in their native ranges (Lundgren et al., 2018); 22% of all introduced mammals are threatened in their native ranges (Lundgren et al., 2024b); 27% of introduced plants are threatened in at least part of their native range (Staude et al., 2025); and many birds and herpetofauna have found refuge through introductions (Gibson and Yong, 2017; Figure 1). What if conservationists were to value at least some of these introduced populations of threatened species?

These paradoxes are not outliers but are increasingly likely to become a core feature of twenty-first century conservation. The world is continuing to change in dramatic ways as humans continue to alter landscapes, global climate and the chemical composition of the environment (Kerr et al., 2025). Are current conservation paradigms sufficient to make pragmatic and conscientious decisions that can prevent extinctions and protect complex and diverse ecosystems?

Including introduced organisms under the umbrella of conservation care is one of the most contentious questions in conservation biology. Yet, we believe it is one of the most essential to wrestle with – on empirical and normative grounds – in order to prepare conservation for a radically novel future (Schlaepfer and Lawler, 2023).

While many of us recently proposed and conducted simulations of how conservation may prevent extinctions by accounting for

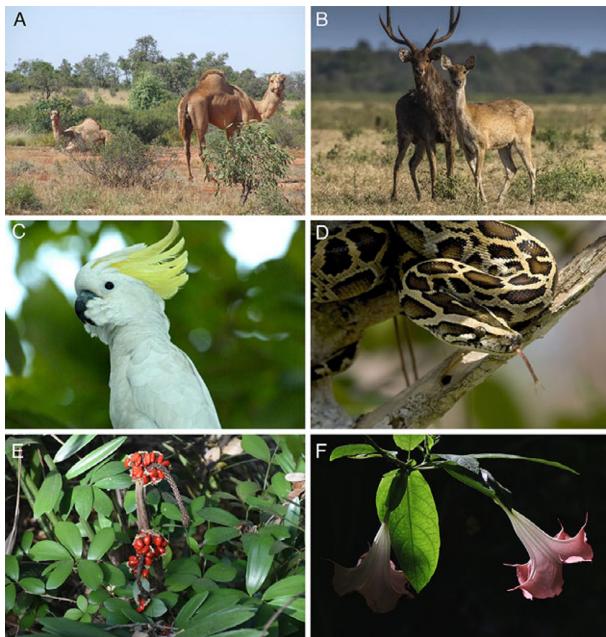


Figure 1. Many introduced organisms are threatened in their native ranges. These organisms present conservation paradoxes that can only be navigated with critical thinking in empirical and normative dimensions. (A) The world’s only population of wild dromedary camels roam in central Australia (extinct in the wild, not listed on the IUCN Red List). (B) Rusa deer (*Rusa timorensis*, vulnerable in their native range) have established populations in eastern Australia (Wallach et al., 2018b); (C) yellow-crested cockatoos (*Cacatua sulphurea*, critically endangered) are thriving in Hong Kong (Andersson, 2023); (D) Burmese pythons (*Python bivittatus*, vulnerable) are considered one of the worst ‘invasive’ species in Florida (IUCN, 2018); (E) cardboard cycad (*Zamia furfuracea*, endangered) is widespread in Florida; and (F) Angel’s trumpet (*Brugmansia suaveolens*) is extinct in the wild but has established wild introduced populations globally. A–F: ©ADW; ©<https://animalia.bio/>; ©Astrid Andersson; ©<https://animalia.bio/>; ©Jens-Christian Svensson; ©Scott Hecker.

introduced populations (Lundgren et al., 2024b), these matters require case-by-case decision-making that considers both global conservation aims (e.g., preventing extinctions) and local conservation concerns (e.g., the effects of the introduced organism). Moreover, as we will describe below, these decisions require attention to both empirical data and normative values for how the world *ought* to be.

We here present a series of questions to help reconcile the contradictions – and opportunities – presented by introduced species in an age of species reshuffling and extinction. These questions are meant to help guide conservationists and ecologists in critically thinking about the effects of introduced organisms and how valuing some introduced populations may assist with global efforts to prevent extinction.

When is nativeness empirically measurable?

The core functional postulate underlying conservation’s concern with the effects of introduced organisms was articulated by Michael Soulé (1985), when he wrote, ‘many of the species that constitute natural communities are the products of coevolutionary processes’. According to this hypothesis, introduced species sever the long-term coevolved relationships between native species, leading to chain reactions that unravel ecosystems. This notion has become central to conservation thought (Rejmánek and Simberloff, 2017; Pauchard et al., 2018).

While there is evidence that some introduced species have prospered, at least on short timescales, because of coevolutionary mismatches (e.g., Shine, 2010; Brian and Catford, 2023), there remains considerable uncertainty regarding the importance of long-term coevolutionary history in shaping ecological interactions. Even specialized interactions can emerge simply from ecological fitting (Janzen, 1985; Wilkinson, 2004); dominant coevolutionary hypotheses in invasion biology have mixed and declining support (Jeschke et al., 2012); and considerable evidence indicates that (co)evolution can occur on rapid, ecologically relevant timescales, even among large vertebrates – suggesting that (co) evolution is a more dynamic force in ecology than typically considered (Carroll et al., 2005; Cattau et al., 2018; Singer and Parmesan, 2018; Campbell-Staton et al., 2021).

However, the primary evidence used to argue for the ‘harmfulness’ of introduced species does not stem from studies of (co) evolutionary mismatches but from the negative effects of introduced species on native species. For instance, the 2023 Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services (IPBES) review of over 30,000 peer-reviewed articles and copious gray literature found that 85% of the effects of introduced species were negative for native species (Roy et al., 2024). At first glance, numbers like this seem indisputable. However, these numbers can be easily misinterpreted because they lack a proper null comparison: the effects of similar native species (Sagoff, 2020).

Ecosystems are built from ‘negative’ interactions: Native predators reduce the abundance of their prey, native herbivores reduce the abundance of their preferred plants, native plants take up space that could have been used by other native plants and so on (Estes et al., 2011). Finding that introduced species reduce the abundance of their food sources or of their competitors, or take up space, only proves that they are alive.

To understand whether the effects of introduced organisms are due to their nativeness *per se*, one must compare their effects to similar native organisms, while controlling for relevant confounding variables (Figure 2). In other words, could an extraterrestrial

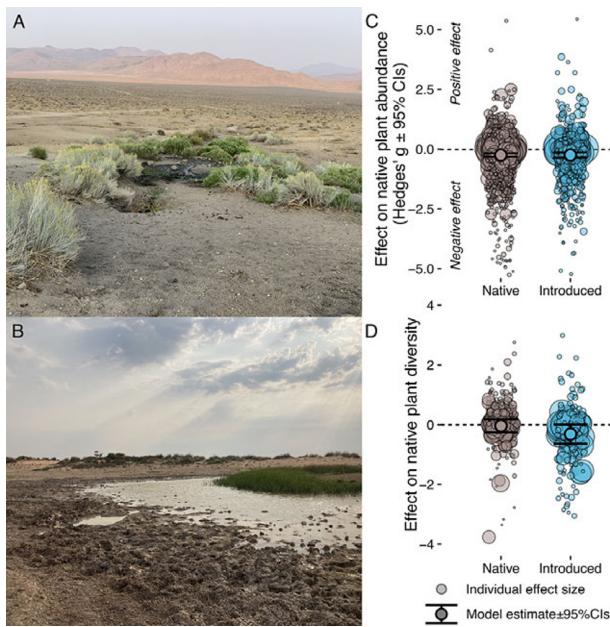


Figure 2. When is nativeness biologically measurable? (a and b) Could an extraterrestrial ecologist empirically determine that the megafauna effects in (a) were caused by introduced megafauna while those in (b) were caused by native megafauna? (a) ©EJL, Feral donkey impacts in Death Valley National Park, (b) ©EJL, native megafauna impacts in the Kalahari, South Africa. (c and d) It does not appear that an extraterrestrial could: A recent systematic meta-analysis of 221 studies found no evidence for differences between the effects of native and introduced megafauna on native plant abundance (c) or diversity (d), with functional traits such as dietary selectivity and body mass instead explaining the effects of native and introduced megafauna alike (Figure from Lundgren et al., 2024a).

ecologist determine which species was native and which introduced from their measurable impacts alone (Brown and Sax, 2005)? Without a proper null comparison, tabulations of the effects of introduced species are insufficient to make any inference about the importance of (non-)nativeness itself and only serve as evidence that the introduced organism exists. As such, tabulations of the impacts of introduced species – such as the International Union for the Conservation of Nature’s (IUCN) environmental impact classification for alien taxa protocol (Hawkins et al., 2015) – present unfalsifiable tautologies where there is no way for an introduced organism to not be ‘harmful’.

Indeed, many, if not most, meta-analyses have found no difference between the effects of introduced and native organisms (e.g., Howard et al., 2017; Boltovskoy et al., 2021; Lundgren et al., 2024a; Wooster et al., 2024a). As an illustrative case in point, consider the Micronesian monitor lizards, which were thought to be introduced by Polynesians, until genetic evidence revealed that these lizards are native and endemic – and now endangered by conservation eradication programs (Weijola et al., 2020).

If one cannot empirically determine the nativeness of an organism based on its impacts, the arguments used to eradicate introduced species could justify the eradication of any living organism. We suggest that identifying when and where nativeness is empirically discernible is a prerequisite to invasion biology’s core claims and should be a key research priority.

In many cases, the effects of introduced species are not only tautological but are confounded with other anthropogenic drivers. For example, of the 256 extinct species on the IUCN Red List for which one or more ‘invasive’ species are listed as a threat, just six have an ‘invasive’ plant mentioned as the primary cause of

extinction. All six plants were found on the Seychelles island of Mahé, and the plant invader in question is *Cinnamomum verum*. However, the cinnamon plant was not simply introduced to the island by humans and then ‘out-competed’ the natives due to its lack of coevolutionary history. Mahé’s forest was almost completely razed by the French for timber in the 1820s, and cinnamon thrived in the starkly novel conditions of the clear-cut (Kueffer, 2013). It appears to us that the primary cause of these extinctions should properly be described as ‘deforestation’. Moreover, if taking up a high proportion of available space on the local scale is enough to be deemed ‘harmful’, would conservation consider old-growth redwoods, mangroves and boreal spruce to also be ‘harmful’?

This is not to say that introduced organisms do not come into conflict with conservation goals. Yet, so do native species (e.g., Bond and Loffell, 2001; Barrett and Stiling, 2006). For example, both native and introduced species can attain high densities and thus have strong effects because of underlying anthropogenic impacts, such as predator persecution, changes in disturbance regimes, nutrient loading, deforestation and so on (MacDougall and Turkington, 2005; Stromberg et al., 2007; King and Tschinkel, 2008; Vasquez et al., 2008; Wallach et al., 2015; Jeppesen et al., 2025). Focusing on eradication or control in these situations distracts us from addressing ultimate drivers and is likely to be ineffective, especially if eradication only leads to the establishment of other species adapted to the novel conditions (Byers, 2002; Kueffer, 2013).

The work of the conservation scientist, in our view, is to determine the ultimate drivers of conservation conflicts with introduced species, which may include coevolutionary mismatches but can also include purely ecological drivers. We suggest the following questions that we believe are foundational to critical thinking about the effects of introduced organisms and can help reveal new opportunities to prevent extinctions and to focus conservation energies in pragmatic directions.

1. Does the introduced organism have effects dissimilar to the effects of similar native species? Or, in other words, could one tell if the organism was introduced if one did not already know?
2. Is there any way for the introduced organism to not be ‘harmful’? – that is, are the claims falsifiable? For example, if a study reports on the effects of introduced species on biodiversity but defines biodiversity as only constituting native species, then the claims are tautological and unfalsifiable.
3. What are the ultimate drivers of the abundance and impacts of an introduced species? Could they be a function of ecological drivers, such as the persecution of predators, nutrient pollution, climate change, changes in disturbance regimes, fire suppression, deforestation and so on?

How ought the world to be?

Normative values are essential in applied scientific disciplines, such as conservation or medicine. Without values, there is no way to decide what one *ought* to do, nor where to direct research attention. Much as medicine is driven by a plurality of values (e.g., the life of the patient, their quality of life, avoiding unnecessary pain and so on), which are sometimes aligned and sometimes in conflict, conservation is also driven by a plurality of values for how the world *ought* to be (Sandbrook et al., 2011). Among these values, *biological nativism* underlies the way many conservation biologists understand introduced organisms and is central to many conservation policies and

treaties (e.g., COP15 Rohwer and Marrs, 2021; Conference of Parties to the UN Convention on Biological Diversity, 2022).

Biological nativism is the belief that systems should be similar to how they were at some moment in the past. These temporal baselines are generally defined by the first descriptions of European explorers or, in Europe, the dawn of industrialization (Peretti, 1998; Hettinger, 2001; IUCN, 2018). However, others (including some members of this authorship team) have argued that the most biologically appropriate baseline for conservation is the range of conditions of the Miocene-Pleistocene (23 million years ago to 12,000 years ago), when large animals set the context under which most modern terrestrial organisms evolved (Faurby and J-C, 2015; Lundgren et al., 2020; Søndergaard et al., 2025). As such, the inadvertent reintroduction of wild horses in North America or even feral camels in Australia – which share functional similarities to extinct Australian megafauna (Lundgren et al., 2020) – can be described as either ‘degradation’ or ‘restoration’ depending on which baseline we choose.

Regardless of its temporal baseline, there are legitimate defenses for nativism as a guiding value in conservation. Nativism can be a posture of respect for the world as it once was, which can counteract the ‘overweening arrogance’ (Gould, 1998) of working to transform ecosystems for esthetic or economic reasons. In this way, nativism can be defended as a form of intellectual humility: treating the past as a guide given our limited understanding of how ecosystems work.

However, nativism can also perpetuate what many consider to be hubristic and ineffective actions, such as attempting to constrain the inherent dynamism of ecosystems and focusing on eradication instead of addressing the fundamental ecological drivers of undesirable outcomes (Gurevitch and Padilla, 2004; Wallach et al., 2018a). In this way, nativism can come into stark conflict with other important values that are shared by conservation practitioners and the public – who fund and support the very existence of conservation. These values include preferences for wildness (i.e., for organisms and communities to be autonomous and self-willed [Ridder, 2007]); the intrinsic value of biological collectives (e.g., species regardless of nativeness); the intrinsic value of sentient nonhuman individuals (Wallach et al., 2018a); and the utilitarian value of services rendered to humans (Sandbrook et al., 2011).

Prioritizing any one of these values over all others as a ‘moral truth’ oversteps the power of science and excludes those with different values. Moreover, pretending that our values are empirical facts is a form of ‘stealth advocacy’ (Cardou and Vellend, 2023) – a major driver of distrust in science with serious consequences to the political legitimacy of conservation and ecology.

While there is no empirical answer for what one *ought* to do that does not involve values or preferences, working toward value transparency and the clear separation of empirical claims from normative claims can help us navigate these complex paradoxes. We thus suggest the following questions to help guide critical thought about how to respond to the paradoxes presented by introduced species and novel ecosystems:

1. Which conservation claims are values in empirical clothes? Terms like ‘ecological health’, ‘invasion’, ‘ecological harm’, ‘pristine’, ‘equilibrium’, ‘degrade’ and ‘disrupt’ are explicitly normative or have deep normative roots. When terms like these are used, even in what seems to be purely empirical research, what values are being deployed?
2. As an exercise of imagination, what would happen if a different set of values were prioritized in a conservation project or

scientific manuscript? How might our priorities and interpretations change?

3. Which values are in conflict or in alignment in a conservation project, and which values are most defensible and aligned with local communities? We suggest that mapping the alignments and conflicts between values in conservation projects can provide insight into an array of pathways for conservation action.

What might a twenty-first century conservation look like?

An evolutionary heartbeat ago, the world was populated by giants whose extinctions led to radical ecological changes (Svenning et al., 2024). More recently, humans have again reshaped the world in a great and ongoing species reshuffling. Many of these introduced species are threatened in their native ranges, and all of them are evolving – a process which will increase global species diversity (Singer and Parmesan, 2018; Faurby et al., 2022). While de-extinction may perhaps create functional analogues of lost species, broadening our conservation ethos can also prevent extinctions – instantly and for free – but requires a seismic shift in conservation thinking.

This shift will allow us to see and consider the invisible biodiversity of introduced organisms and the unexpected echoes of prehistoric ecologies in novel ecosystems (Wallach et al., 2018b; Wooster et al., 2024a). Doing so requires attention to appropriate scientific comparisons, as well as explicit transparency about our values, critical justification of those values and not mistaking values for empirical facts. This will not provide easy answers, but is necessary to face an increasingly novel future (Ordonez et al., 2024). After all, rapid changes in global climate and land use are likely to scramble the ranges of all species into never-before-seen configurations (Ordonez et al., 2024; Kerr et al., 2025).

To face these challenges, we suggest that twenty-first century conservation will need to expand its vision of what is possible and what is good. This does not mean that we abandon our roles as stewards, who intervene to prevent extinctions or other agreed-upon ecological losses. Instead, we should critically examine circular claims regarding the harmfulness of introduced organisms as we work to identify the actions that can best conserve planetary-scale biodiversity, even as it kaleidoscopes into novel configurations. As such, we believe that twenty-first century conservation would benefit from embracing a pluralistic and future-facing ethos that is inspired by many pasts, and that is transparent to the diversity of values that undergird our love for the more-than-human world.

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Data availability statement. This article did not conduct quantitative analyses or use data. The data plotted in Figure 2 is from Lundgren et al. (2024a).

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References

Andersson ALA (2023) Ongoing and emergent threats to yellow-crested cockatoos (*Cacatua sulphurea*): a critically endangered species surviving in a city. HKU Theses Online (HKUTO). The University of Hong Kong (Pokfulam, Hong Kong)

Barrett MA and Stiling P (2006) Impacts of endangered key deer herbivory on imperiled pine rocklands: A conservation dilemma? *Animal Biodiversity and Conservation* 29(2), 165–178.

Boltovscky D, Correa NM, Burlakova LE, Karatayev AY, Thuesen EV, Sylvester F and Paolucci EM (2021) Traits and impacts of introduced species: A quantitative review of meta-analyses. *Hydrobiologia* 848(9), 2225–2258.

Bond WJ and Loffeld D (2001) Introduction of giraffe changes acacia distribution in a south African savanna. *African Journal of Ecology* 39(3), 286–294.

Brian JI and Catford JA (2023) A mechanistic framework of enemy release. *Ecology Letters* 26(12), 2147–2166.

Brown JH and Sax DF (2005) Biological invasions and scientific objectivity: Reply to Cassey *et al.* (2005). *Austral Ecology* 30(4), 481–483. (accessed 12 March 2025).

Byers JE (2002) Impact of non-indigenous species on natives enhanced by anthropogenic alteration of selection regimes. *Oikos* (Copenhagen, Denmark) 97(3), 449–458.

Campbell-Staton SC, Arnold BJ, Gonçalves D, Granli P, Poole J, Long RA and Pringle RM (2021) Ivory poaching and the rapid evolution of tusklessness in African elephants. *Science (New York, N.Y.)* 374(6566), 483–487.

Cardou F and Vellend M (2023) Stealth advocacy in ecology and conservation biology. *Biological Conservation* 280, 109968.

Carroll SP, Loya JE, Dingle H, Mathieson M, Famula TR and Zalucki MP (2005) And the beak shall inherit - evolution in response to invasion. *Ecology Letters* 8(9), 944–951.

Cattau CE, Fletcher RJ Jr, Kimball RT, Miller CW and Kitchens WM (2018) Rapid morphological change of a top predator with the invasion of a novel prey. *Nature Ecology & Evolution* 2(1), 108–115.

Conference of Parties to the UN Convention on Biological Diversity (2022) Kunming-Montreal Global Biodiversity Framework CBD/COP/15/L25. <https://www.cbd.int/conferences/2021-2022/cop-15/documents>

Estes JA, Terborgh J, Brashares JS, Power ME, Berger J, Bond WJ, Carpenter SR, Essington TE, Holt RD, Jackson JBC, Marquis RJ, Oksanen L, Oksanen T, Paine RT, Pikitch EK, Ripple WJ, Sandin SA, Scheffer M, Schoener TW, Shurin JB, Sinclair ARE, Soulé ME, Virtanen R and Wardle DA (2011) Trophic downgrading of planet earth. *Science* 333(6040), 301–306.

Faurby S, Davis M, Pedersen RØ, Schowanek SD, Antonelli A and Svenning J-C (2018) PHYLACINE 1.2: The phylogenetic atlas of mammal macroecology. *Ecology* 99(11), 2626.

Faurby S, Pedersen RØ, Svenning J-C and Antonelli A (2022) The counteracting effects of anthropogenic speciation and extinction on mammal species richness and phylogenetic diversity. *Global Ecology and Biogeography: A Journal of Macroecology* 31(9), 1810–1823.

Faurby S and Svenning J-C (2015) Historic and prehistoric human-driven extinctions have reshaped global mammal diversity patterns. *Diversity & Distributions* 21(10), 1155–1166.

Gibson L and Yong DL (2017) Saving two birds with one stone: Solving the quandary of introduced, threatened species. *Frontiers in Ecology and the Environment* 15(1), 35–41.

Gould SJ (1998) An evolutionary perspective on strengths, fallacies, and confusions in the concept of native plants. *Arnoldia Zimbabwe* 58(1), 11–19.

Gurevitch J and Padilla DK (2004) Are invasive species a major cause of extinctions? *Trends in Ecology & Evolution* 19(9), 470–474.

Hawkins CL, Bacher S, Essl F, Hulme PE, Jeschke JM, Kühn I, Kumschick S, Nentwig W, Pergl J, Pyšek P, Rabitsch W, Richardson DM, Vilà M, Wilson JRU, Genovesi P and Blackburn TM (2015) Framework and guidelines for implementing the proposed IUCN environmental impact classification for alien taxa (EICAT). *Diversity & Distributions* 21(11), 1360–1363. (accessed 12 March 2025).

Hettinger N (2001) Exotic species, naturalisation, and biological nativism. *Environmental Values* 10(2), 193–224.

Howard BR, Therriault TW and Côté IM (2017) Contrasting ecological impacts of native and non-native marine crabs: A global meta-analysis. *Marine Ecology Progress Series* 577, 93–103.

IUCN (2018) IUCN Red List. <https://www.iucnredlist.org/> (accessed 17 June 2023).

Janzen DH (1985) On ecological fitting. *Oikos* 45(3), 308–310.

Jeppesen R, de Rivera CE, Grosholz ED, Tinker MT, Hughes BB, Eby R and Wasson K (2025) Recovering population of the southern sea otter suppresses a global marine invader. *Biological Invasions* 27(1), 1–13.

Jeschke JM, Gómez Aparicio L, Haider S, Heger T, Lortie CJ, Pyšek P and Strayer DL (2012) Support for major hypotheses in invasion biology is uneven and declining. *NeoBiota* 14, 1–20.

Kerr MR, Ordonez A, Riede F, Atkinson J, Pearce EA, Sykut M, Trepel J and Svenning J-C (2025) Widespread ecological novelty across the terrestrial biosphere. *Nature Ecology & Evolution* 1–10. (accessed 27 March 2025).

King JR and Tschinkel WR (2008) Experimental evidence that human impacts drive fire ant invasions and ecological change. *Proceedings of the National Academy of Sciences of the United States of America* 105(51), 20339–20343.

Kueffer C (2013) Integrating natural and social sciences for understanding and managing plant invasions. *Biodiversity and Society in the Pacific Islands* 71–96.

Lundgren EJ, Bergman J, Trepel J and le Roux E, Monsarrat S, Kristensen JA, Pedersen RØ, Pereyra P, Tietje M, Svenning JC (2024a) Functional traits—Not nativeness—Shape the effects of large mammalian herbivores on plant communities. *Science* 383(6682), 531–537. <https://doi.org/10.1126/science.adh2616>.

Lundgren EJ, Ramp D, Ripple WJ and Wallach AD (2018) Introduced megafauna are rewilding the Anthropocene. *Ecography* 41, 857–866.

Lundgren EJ, Ramp D, Rowan J, Middleton O, Schowanek SD, Sanisidro O, Carroll SP, Davis M, Sandom CJ, Svenning J-C and Wallach AD (2020) Introduced herbivores restore late Pleistocene ecological functions. *Proceedings of the National Academy of Sciences of the United States of America* 117(14), 7871–7878.

Lundgren EJ, Wallach AD, Svenning J-C, Schlaepfer MA, Andersson ALA and Ramp D (2024b) Preventing extinction in an age of species migration and planetary change. *Conservation Biology* e14270. <https://doi.org/10.1101/2023.10.17.562809>.

MacDougall AS and Turkington R (2005) Are invasive species the drivers or passengers of change in degraded ecosystems? *Ecology* 86(1), 42–55. (accessed 14 December 2024).

Ordonez A, Riede F, Normand S and Svenning J-C (2024) Towards a novel biosphere in 2300: Rapid and extensive global and biome-wide climatic novelty in the Anthropocene. *Philosophical Transactions of the Royal Society of London. Series B, Biological Sciences* 379(1902), 20230022.

Pauchard A, Meyerson LA, Bacher S, Blackburn TM, Brundu G, Cadotte MW, Courchamp F, Essl F, Genovesi P, Haider S, Holmes ND, Hulme PE, Jeschke JM, Lockwood JL, Novoa A, Nuñez MA, Peltzer DA, Pyšek P, Richardson DM, Simberloff D, Smith K, van Wilgen BW, Vilà M, Wilson JRU, Winter M and Zenni RD (2018) Biodiversity assessments: Origin matters. *PLoS Biology* 16(11), e2006686.

Peretti JH (1998) Nativism and nature: Rethinking biological invasion. *Environmental Values* 7(2), 183–192.

Price GJ, Ferguson KJ, Webb GE, Feng Y-X, Higgins P, Nguyen AD, Zhao J-X, Joannes-Boyau R and Louys J (2017) Seasonal migration of marsupial megafauna in Pleistocene Sahul (Australia-New Guinea). *Proceedings. Biological sciences* 284(1863). <https://doi.org/10.1098/rspb.2017.0785>.

Prideaux GJ, Ayliffe LK, DeSantis LRG, Schubert BW, Murray PF, Gagan MK and Cerling TE (2009) Extinction implications of achenopod browse diet for a giant Pleistocene kangaroo. *Proceedings of the National Academy of Sciences of the United States of America* 106(28), 11646–11650.

Rejmánek M and Simberloff D (2017) Origin matters. *Environmental Conservation* 44(2), 97–99. (accessed 21 June 2023).

Ridder B (2007) The naturalness versus wildness debate: Ambiguity, inconsistency, and unattainable objectivity. *Restoration Ecology* 15(1), 8–12. (accessed 21 March 2025).

Rohwer Y and Marris E (2021) Ecosystem integrity is neither real nor valuable. *Conservation Science and Practice* 3(4). <https://doi.org/10.1111/csp2.411>.

Roy HE, Pauchard A, Stoett P and Renard Truong T (2024) IPBES Invasive Alien Species Assessment: Full report. Zenodo. <https://doi.org/10.5281/ZENODO.11629357>.

Sagoff M (2020) Fact and value in invasion biology. *Conservation Biology: The Journal of the Society for Conservation Biology* 34(3), 581–588.

Sandbrook C, Scales IR, Viri B and Adams WM (2011) Value plurality among conservation professionals. *Conservation Biology: The Journal of the Society for Conservation Biology* 25(2), 285–294.

Schlaepfer MA and Lawler JJ (2023) Conserving biodiversity in the face of rapid climate change requires a shift in priorities. *WIREs Climate Change* 14(1). <https://doi.org/10.1002/wcc.798>.

Shine R (2010) The ecological impact of invasive cane toads (*Bufo marinus*) in Australia. *The Quarterly Review of Biology* 85(3), 253–291.

Singer M and Parmesan C (2018) Lethal trap created by adaptive evolutionary response to an exotic resource. *Nature* 557, 238–241.

Søndergaard SA, Fløjgaard C, Ejrnæs R and Svenning J-C (2025) *Shifting baselines and the forgotten giants: Integrating megafauna into plant community ecology*. Copenhagen, Denmark: Oikos, e11134. (accessed 27 March 2025).

Soulé ME (1985) What is conservation biology? *Bioscience* 35(11), 727–734.

Staude IR, Grenié M, Thomas CD, Kühn I, Zizka A, Golivets M, Ledger SEH and Méndez L (2025) Many non-native plant species are threatened in parts of their native range. *The New Phytologist*. <https://doi.org/10.1111/nph.70193>.

Stromberg JC, Lite SJ, Marler R, Paradzick C, Shafroth PB, Shorrock D, White JM and White MS (2007) Altered stream-flow regimes and invasive plant species: The Tamarix case. *Global Ecology and Biogeography: A Journal of Macroecology* 16(3), 381–393.

Svenning JC, Lemoine RT and Bergman J, Buitenhof R, le Roux E, Lundgren E, Mungi N, Pedersen RØ (2024) *The late-Quaternary megafauna extinctions: patterns, causes, ecological consequences, and implications for ecosystem management in the Anthropocene*. Cambridge Prisms 2, e5. <https://doi.org/10.1017/ext.2024.4>.

Vasquez E, Sheley R and Svejcar T (2008) Nitrogen enhances the competitive ability of cheatgrass (*Bromus tectorum*) relative to native grasses. *Invasive Plant Science and Management* 1(3), 287–295.

Wallach AD, Bekoff M, Batavia C, Nelson MP and Ramp D (2018a) Summoning compassion to address the challenges of conservation. *Conservation Biology: The Journal of the Society for Conservation Biology* 32(6), 1255–1265.

Wallach AD, Lundgren EJ, Ripple WJ and Ramp D (2018b) Invisible megafauna. *Conservation Biology: The Journal of the Society for Conservation Biology* 32(4), 962–965.

Wallach AD, Ripple WJ and Carroll SP (2015) Novel trophic cascades: Apex predators enable coexistence. *Trends in Ecology & Evolution* 30(3), 146–153.

Weijola V, Vahtera V, Koch A, Schmitz A and Kraus F (2020) Taxonomy of Micronesian monitors (Reptilia: Squamata: Varanus): Endemic status of new species argues for caution in pursuing eradication plans. *Royal Society Open Science* 7(5), 200092.

Wilkinson DM (2004) The parable of Green Mountain: Ascension Island, ecosystem construction and ecological fitting. *Journal of Biogeography* 31(1), 1–4.

Wooster EIF, Middleton OS, Wallach AD, Ramp D, Sanisidro O, Harris VK, Rowan J, Schowanek SD, Gordon CE, Svenning J-C, Davis M, Scharlemann JPW, Nimmo DG, Lundgren EJ and Sandom CJ (2024a) Australia's recently established predators restore complexity to food webs simplified by extinction. *Current Biology: CB* 34(22), 5164–5172, e2.

Wooster E, Ramp D, Lundgren EJ, Bonsen GT, Geisler-Edge A, Ben-Ami D, Carthey A, Carroll S, Keynan O, Olek Y, O'Neill A, Shanas U and Wallach AD (2024b) Prey responses to foxes are not determined by nativeness. *Ecography* e07031. <https://doi.org/10.1111/ecog.07031>.