

1 **Quantifying seasonal to multi-decadal signals in coastal water quality using**  
2 **high- and low-frequency time series data**

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11  
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14  
15 **Abstract**

16 To inform water quality monitoring techniques and modelling at coastal research sites,  
17 this study investigated seasonality and trends in coastal lagoons on the eastern shore of Virginia,  
18 USA. Seasonality was quantified with harmonic analysis of low-frequency time-series,  
19 approximately 30 years of quarterly sampled data at thirteen mainland, lagoon, and ocean inlet  
20 sites, along with 4-6 years of high-frequency, 15-minute resolution sonde data at two mainland  
21 sites. Temperature, dissolved oxygen, and apparent oxygen utilization (AOU) seasonality were  
22 dominated by annual harmonics while salinity and chlorophyll-a exhibited mixed annual and  
23 semi-annual harmonics. Mainland sites had larger seasonal amplitudes and higher peak summer  
24 values for temperature, chlorophyll-a and AOU, likely from longer water residence times,  
25 shallower waters, and proximity to marshes and uplands. Based on statistical subsampling of  
26 high-frequency data, one to several decades of low-frequency data (at quarterly sampling) were

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27 needed to quantify the climatological seasonal cycle within specified confidence intervals.  
28 Statistically significant decadal warming and increasing chlorophyll-a concentrations were found  
29 at a sub-set of mainland sites, with no distinct geographic patterns for other water quality trends.  
30 The analysis highlighted challenges in detecting long-term trends in coastal water quality at sites  
31 sampled at low frequency with large seasonal and interannual variability.

32

### 33 **Impact Statement**

34 Accurate monitoring of interconnected water quality variables such as temperature,  
35 salinity, dissolved oxygen, chlorophyll-a, and apparent oxygen utilization (AOU) is vital for  
36 tracking the status, health, and dynamics of coastal marine ecosystems. For example, warming  
37 water temperatures can enhance the frequency and severity of algal blooms (elevated levels of  
38 chlorophyll), leading to the formation of hypoxic and anoxic zones (low to no dissolved oxygen),  
39 and influencing the community metabolism (AOU which depends on a balance of  
40 photosynthesis, respiration, and ventilation). However, there are logistical and scientific  
41 challenges in maintaining consistent and adequate water quality monitoring. High-frequency *in-*  
42 *situ* measurements using automated sondes have a high temporal frequency (i.e., every 15  
43 minutes) and are less labor intensive. However, due to power and maintenance demands these  
44 are mostly confined to shore-based sites or more geographically limited and expensive coastal  
45 scientific moorings and cabled arrays. Automated instruments are also subject to sensor  
46 malfunction or biofouling leading to a consequent loss of consistent time series without frequent  
47 upkeep. Longer term sites measured at a low-frequency have higher geographic spread and are  
48 sampled using a manual sonde, water sampling and lab extraction methods typically at a lower  
49 temporal frequency (i.e., weekly to quarterly), but are limited to safe conditions for shore-based

50 or boat sampling. In this study, we investigated if there are differences between seasonal  
51 harmonic elements of high- and low-frequency data over site types and how sampling frequency  
52 affected estimates of the magnitude and timing of the seasonal cycle using sub-sampling, i.e.  
53 sites measured at a high-frequency subsampled at the rate of sites measured at a low-frequency.  
54 Additionally using boot-strapping techniques, we explored how many years of simulated  
55 quarterly sampling would be needed to quantify the climatological seasonal cycle within  
56 specified confidence intervals. We hope that this research can guide water quality monitoring  
57 techniques and modeling at other coastal research sites in order to be able to adequately observe  
58 ecosystems in a changing climate.

59

## 60 **Introduction**

61 Coastal lagoons are typically shallow with submerged, often vegetated flats and  
62 intermittent openings to the sea (Scanes et al., 2007). The marshes, oyster reefs, and seagrass  
63 meadows in temperate coastal lagoons provide vital ecosystem services including biodiversity,  
64 shoreline protection, and carbon storage (Smith et al., 2022; Orth et al., 2020). These systems are  
65 vulnerable to environmental change, including habitat loss from coastal development,  
66 degradation of water quality from development and pollution, changes in biodiversity due to  
67 overexploitation, and invasive species, and climate change impacts, especially marine heat waves  
68 and sea-level rise (Newton et al., 2018). Water quality variables, such as temperature, salinity,  
69 dissolved oxygen (DO), apparent oxygen utilization (AOU), and chlorophyll-a (Chl) integrate  
70 responses to human perturbation and climate change and are monitored to document long-term  
71 change.

72           The Eastern Shore of coastal Virginia, USA, has an extensive coastal lagoon and barrier  
73 island system marked by a large expanse of relatively undeveloped rural coastline and lacks  
74 significant inputs of fluvial sources of freshwater and sediment (Safak et al., 2015). The lagoons  
75 studied here are on the eastern side of the Delmarva Peninsula fronting the coastal ocean in the  
76 Mid-Atlantic Bight, with no direct hydrologic connection to the Chesapeake Bay estuary on the  
77 western side of the Peninsula. Historically, this area, the Virginia Coast Reserve (VCR), hosted  
78 large scallop fisheries due to the abundance of *Zostera marina* seagrass beds until a massive  
79 seagrass die-off event in the 1930s (Hondula and Pace, 2014; Oreska et al., 2017). A large-scale  
80 restoration effort by reseeded that began in 2001, has returned over 36 km<sup>2</sup> of seagrass as of  
81 2021 (Orth et al., 2020; Oreska et al., 2021).

82           In the VCR coastal lagoons, water quality varies as a function of season, tidal currents  
83 and flushing, winds, and storm conditions (Hondula and Pace, 2014). Climate is the most  
84 dominant driver of ecological change, especially sea-level rise, storms, increased temperature  
85 and marine heat waves (McGlathery et al., 2013). The VCR coastal lagoons have relatively low  
86 nutrient loading and water column chlorophyll concentrations (Carr et al., 2012; McGlathery et  
87 al., 2007), with little variation in salinity from adjacent coastal ocean waters due to limited  
88 freshwater discharge (Oreska et al., 2021). This good water quality largely reflects the low  
89 fluvial inputs, coastal development, and human influence for the VCR coastal lagoons. This  
90 differs from the more impaired estuary systems typical along the Mid-Atlantic seaboard,  
91 including the Chesapeake Bay, where substantial point source and non-point nutrient source  
92 pollution causes extensive coastal eutrophication and low-oxygen conditions (Sabo et al., 2022).  
93 The VCR coastal lagoons thus can serve, more generally, as an end-member for the pre-

94 industrial low-disturbance or future recovery state of coastal water quality at other temperate  
95 monitoring and research sites.

96 The Virginia Coast Reserve Long-Term Ecological Research (VCR-LTER) project  
97 maintains thirteen long-term (multi-decade) water quality monitoring sites measured at a low-  
98 frequency within the coastal lagoon system. The sampling network spans across environment  
99 types, water depths, and residence times with respect to tidal flushing (Safak et al., 2018). Most  
100 of the VCR-LTER sites are relatively shallow and oligotrophic, with a mean semi-diurnal tidal  
101 range of 1.2 m (Oreska et al., 2021). The Virginia Institute of Marine Science (VIMS) at the  
102 Eastern Shore Laboratory (ESL) (referred to as ESL hereafter) recently established two water  
103 quality monitoring sites with high-frequency, automated sonde measurements; both sites are  
104 shore-based on the mainland side of the lagoon system.

105 Data analysis of seasonal patterns in multi-year water quality data with persistent gaps  
106 can be accomplished using harmonic analysis, a method that represents fluctuations in a time  
107 series from the sum of sine and cosine functions that have different frequencies (Wilks, 2011).  
108 Higher harmonics indicate higher frequencies, with the first harmonic representing in this study  
109 one full annual cycle and the second harmonic representing the semi-annual cycle (Wilks, 2011).  
110 The relative importance of each of these harmonics can be quantified using their respective  
111 fractions of total variance captured in the harmonic analysis, where for most variables the  
112 majority of variance is captured by the first one or two harmonics. This study focused on the first  
113 and second annual harmonics because they contained a majority of the variances. A composite  
114 harmonic, the sum of the first and second harmonics, allows an analysis of the relative  
115 importance of each harmonic in the resulting fitted curve and harmonic model elements.  
116 Deviations from fitted harmonic curves can indicate anomalies from the climatological seasonal

117 cycle, deseasonalized data, which allow for documentation of long-term trends and sub-seasonal  
118 variability. Seasonal harmonic elements such as amplitude, phase shift and minimum/maximum  
119 values can reveal important information about seasonal cycles. Below we use the VCR coastal  
120 lagoon data to illustrate the utility of harmonic seasonal analysis and data sub-sampling  
121 techniques to coastal water quality monitoring sites.

122

## 123 **Materials and Methods**

### 124 *Study Area and Data Description*

125         The study was conducted with VCR-LTER data from the Eastern Shore of Virginia, USA  
126 along the mid-latitude western continental boundary of the North Atlantic (Figure 1). High-  
127 frequency ESL data (temperature, salinity, dissolved oxygen and chlorophyll-a) (Ross and  
128 Snyder, 2023) were measured every 15 minutes using YSI EXO2 Multiparameter Sondes  
129 attached to dual land-based pumps at two creek sites Wachapreague (W) and Willis Wharf  
130 (WW); the dual pump intakes were switched and cleaned weekly, the flow cell wall lightly  
131 cleaned monthly, and the datasondes calibrated every 90 days using a calibration standard and  
132 KorEXO software (Ross & Snyder 2023; Figure 1). The data range from March 25th, 2016 to  
133 December 31st, 2022 at Wachapreague and October 12th, 2018 to December 24th, 2022 at Willis  
134 Wharf, with large gaps in both data (Ross and Snyder, 2023; Table 1). As part of standard data  
135 quality assessment-quality control (QA/QC) methods, suspicious data or outliers were removed,  
136 excluding points that were outside of  $\pm 1$  standard deviation (derived from yearly statistics of  
137 raw data) from the preceding data point (Ross & Snyder 2023).

138         Low-frequency VCR-LTER data (temperature, salinity, and dissolved oxygen), spanning  
139 20-30 years, (McGlathery and Christian, 2022) were collected manually using a YSI Datasonde

140 lowered from a small boat or shore at each of the 13 sites; the datasondes were calibrated  
141 quarterly to annually following the manufacturer recommended procedure using calibration  
142 standards. Discrete water samples (200 ml) collected for chlorophyll-a were filtered through  
143 Whatman GF/F filters (0.7  $\mu\text{m}$  pore size), extracted in the dark for 24 hr (90% acetone), and  
144 concentrations measured using a bench-top Shimadzu 1280 spectrophotometer (McGlathery and  
145 Christian, 2022; Table 1). Minimal QA/QC methods were applied to the database version of  
146 VCR LTER data to remove obvious outliers and bad data points. Water residence times with  
147 respect to tidal flushing were estimated using a three-dimensional, finite-volume coastal ocean  
148 model (FVCOM) that was then validated with field observations (Safak et al., 2018). Before  
149 conducting the harmonic and trend analysis, an additional data screening was conducted on any  
150 remaining outliers in both the ESL and VCR data sets using Chauvenet's Criterion (Glover et al.,  
151 2011).

152 The base-10 logarithm ( $\log_{10}$ ) of chlorophyll-a was taken in order to make the  
153 chlorophyll data more closely normally distributed following typical practice for marine bio-  
154 optical data (Campbell, 1995). AOU was derived from temperature, salinity, and dissolved  
155 oxygen measurements and was calculated using Matlab code by Peltzer (2013) based on  
156 equations from Garcia and Gordon (1992). Because we lacked sufficient resolution of diurnal  
157 variability of photosynthesis and respiration for the sites of low-frequency data collection, no  
158 attempt was made to correct for possible aliasing of the diurnal cycle into VCR AOU estimates;  
159 daily averaging removed the diurnal cycle from the high-frequency ESL data.

#### 160 *Data Analysis*

161 The water quality data were analyzed using a combination of harmonic analysis and  
162 linear regression with minimal *a priori* assumptions. Namely the approach requires that a

163 substantial fraction of the seasonal variability is captured by the annual and semi-annual  
 164 harmonics and that residuals after removing the harmonics exhibit gradual trends that be  
 165 captured by a linear function (versus a step function, quadratic, etc.); both assumptions hold  
 166 reasonably well as shown in the Results and Discussion section.

167 Data for each site were compiled into climatological day of year graphs, and a harmonic  
 168 model fit estimate,  $y^m(t)$ , was constructed using 1<sup>st</sup> and 2<sup>nd</sup> seasonal harmonics (sine curves) fit  
 169 (Wilks, 2011):

$$170 \quad y^m(t) = \bar{y} + \sum_k^2 \left( a_k \cos \left[ \frac{2\pi kt}{n} \right] + b_k \sin \left[ \frac{2\pi kt}{n} \right] \right) \quad (1)$$

171 where  $k$  was the respective harmonic,  $\bar{y}$  was the mean of the  $y$  values,  $a_k$  and  $b_k$  were the  
 172 cosine and sine coefficients of the  $k$ th harmonic respectively,  $t$  was day of year,  $n$  was annual  
 173 time period (365 days). Harmonic amplitude ( $a_k$ ), phase shift ( $\phi_k$ ), and date of maximum  
 174 values ( $t_{max}$ ) were calculated using methods by Wilks (2011). Variances from harmonic fit ( $\sigma_k^2$ )  
 175 and total fitted variance ( $\sigma_{fit}^2$ ) were calculated using methods by Burroughs, 2003. Confidence  
 176 intervals and reduced chi squared values ( $\chi^2_{red}$ ) were calculated using methods by Glover et al.  
 177 (2011). Only statistically significant differences in harmonic parameters are highlight in the  
 178 Results and Discussion section.

179 The date of the seasonal minimum ( $t_{min}$ ) and maximum ( $t_{max}$ ) of the composite  
 180 harmonics were computed using the zero-points of the first derivative with respect to time,  $t$ , of  
 181 Equation 1. The amplitude of the composite harmonics was found from:

$$182 \quad \sigma_{fit} = \frac{\sigma_{fit}(t_{max}) - \sigma_{fit}(t_{min})}{2} \quad (2)$$

183 Bootstrapping was used to generate confidence intervals for the model parameters and resulting  
 184 harmonic curves.



185 Generalized Least Squares (GLS) was applied to the deseasonalized data anomalies  
 186 calculated using Equation 3 to see if there were any long-term trends for each of the sites and  
 187 variables measured at low frequency.

$$188 \quad \hat{y} = \hat{\beta} - y^m(t_i) \quad (3)$$

189 The long-term, low-frequency data sets were each broken at their midpoint into two equal  
 190 length halves of 7-15 years, depending on the dataset length, and separate harmonic curves were  
 191 calculated for each half using Equation 1. The differences between the harmonic elements for the  
 192 two time periods were calculated. The statistical significance of geographic and temporal  
 193 differences in mean values were assessed and p-values reported using a Student's t-test (Glover  
 194 et al., 2011).

#### 195 *Simulated Low-Frequency Sampling*

196 The ESL high-frequency time-series were sub-sampled to create data sets with similar  
 197 resolution as the VCR-LTER data to explore trade-offs among sampling frequency, duration, and  
 198 climatological seasonal cycle resolution. The first set of experiments to evaluate the skill of the  
 199 VCR-LTER data to resolve climatological cycles was conducted by randomly subsampling the  
 200 daily averaged high-frequency data at the same sampling frequencies as the VCR-LTER  
 201 mainland sites for each variable. The low-frequency sampling was conducted at a consistent  
 202 point in the tidal cycle with the outgoing tide, and the daily averaging was applied to the high-  
 203 frequency to minimize sub-diurnal tidal variability effects. Harmonic curves with confidence  
 204 intervals were computed for the 200 sub-sampled low-frequency series (trials) for each  
 205 parameter using Equation 2. Root mean square error (RMSE) was computed between the low-  
 206 frequency and high-frequency harmonic fits for each of the 200 trials with Equation 4.

$$RMSE = \sqrt{\frac{\sum_{i=1}^{365} (y_{full}(t_i) - y_{sub}(t_i))^2}{n}} \quad (4)$$

where  $y_{full}(t_i)$  was the  $i$ th year day of the full, high-frequency harmonic fit (predicted value),  $y_{sub}(t_i)$  was the subsampled, low-frequency harmonic fit value of that year day (observed value), and  $n$  was the variable-dependent number of subsampled values in the trial used as an estimate of degrees of freedom. RMSE was normalized (nRMSE) by dividing by the standard deviation of the deseasonalized anomalies from the full high-frequency harmonic data set calculated in Equation 3.

The Nash-Sutcliffe Efficiency Coefficient (NSE) is a metric for comparing goodness of fit across models with a maximum NSE=1 (no error) and NSE=0 indicating that the error variance is comparable to the observed variance (Nash and Sutcliffe, 1970).

$$NSE = 1 - \left(\frac{\sigma_{sub}}{\sigma_{obs}}\right)^2 \quad (5)$$

Where  $\sigma_{sub}$  is the standard deviation of the randomly subsampled values for each trial. NSE values were calculated for each variable to estimate fit of the 200 randomly subsampled low-frequency trials versus the full high-frequency data, and the average was taken. Ritter and Muñoz-Carpena (2012) developed criteria to estimate the goodness of fit using this value, where NSE's above 0.90 are very good, 0.80-0.90 are good, 0.65-0.80 are acceptable, and below 0.65 are unsatisfactory.

A second set of experiments was performed to test the effect of time-series duration on resolving climatological cycles by varying the low-frequency sub-sampling rate for the Wachapreague and Willis Wharf site data sorted into seasons. Simulated time-series were constructed by randomly subsampling the high-frequency data  $n_{sub} = 2 \times n_{obs}$  times per season (where  $n_{obs}$  is the maximum number of years for VCR time series). The

229  $\square_{\square\square\square-\square\square\square\square}$  value can be interpreted as the number of years (duration) of quarterly sampling  
 230 or equivalently the duration multiplied by the quarterly sampling density. For each variable and  
 231  $\square_{\square\square\square-\square\square\square\square}$  value (seasonal sampling density), 200 simulated time-series (trials) were  
 232 generated, harmonic analyses were performed for both the composite harmonic and dominant  
 233 harmonic, and average values computed for each harmonic fit parameter for the climatological  
 234 seasonal cycle. RMSE and nRMSE for low-frequency versus high-frequency were calculated for  
 235 each  $\square_{\square\square\square-\square\square\square\square}$  value using Equation 4 to assess error with  $n_{sub} = 4m_{sub-sample}$ .

236 Minimum sampling duration was determined by finding the lowest  $\square_{\square\square\square-\square\square\square\square}$  value  
 237 where the sub-sampled composite harmonic elements for two sequential  $\square_{\square\square\square-\square\square\square\square}$  values  
 238 fell within the high-frequency confidence intervals. In the special case where these intervals were  
 239 not met, the closest value was selected, and that variable was noted. For the respective dominant  
 240 harmonic, the standard deviations were calculated for the date of maximum and seasonal  
 241 amplitudes from the 200 trials. These values were normalized (nSD) by calculating the effective  
 242 asymptote with respect to high  $\square_{\square\square\square-\square\square\square\square}$  values, the average of the standard deviations  
 243 of the 100th-200th trials.

244

## 245 **Results and Discussion**

246 Coastal water quality metrics provide valuable information for assessing trends in  
 247 response to environmental drivers such as pollution and climate change (Buelo et al., 2024;  
 248 Tassone et al., 2021). In the lagoons of the Virginia Coast Reserve, strong seasonal patterns were  
 249 seen in all water quality variables at both high- and low- frequency sampled sites.

250 *Temperature*

251 For temperature, all sites were dominated by the first harmonic, indicating an annual  
252 climatological cycle primarily driven by meteorological factors and physical hydrological  
253 characteristics (Figure 2; Benyahya et al., 2007; Wiberg, 2023). Mainland sites' statistically  
254 significant earlier dates for summer peak temperatures (average of  $213.6 \pm 1.6$  year day,  
255  $p=0.0005$ ), and statistically significant higher maximum peak values (average of  $29.21 \pm 0.61$   
256  $^{\circ}\text{C}$ ,  $p=0.0005$ ) and seasonal amplitudes (average of  $11.77 \pm 0.08$   $^{\circ}\text{C}$ ,  $p=0.007$ ) were linked to their  
257 shallower depths that allowed for more rapid and intense temperature changes, while ocean inlet  
258 and mid-lagoon sites (average of  $218.3 \pm 2.4$  year day,  $27.51 \pm 0.66$   $^{\circ}\text{C}$ , and  $11.26 \pm 0.30$   $^{\circ}\text{C}$ )  
259 were dampened by routine exposure to cooler ocean water and shorter residence times (Safak et  
260 al., 2018; Supplemental Table S1).

#### 261 *Salinity*

262 For salinity there was a mix of annual and semi-annual harmonics at the sites, differing  
263 by site geography, with yearly variations depending on the effects of mixing, evaporation, run-  
264 off, and precipitation (Sachithanathan, 1969). All of the mainland sites were dominated by the  
265 first harmonic, and most ocean inlet and mid-lagoon sites were dominated by the second  
266 harmonic (Figure 2). In general, ocean inlet and mid-lagoon sites showed less intra-site and  
267 temporal variability due to mitigation of the stronger impact of tidal flushing of the nearby ocean  
268 which had an overall more stable seasonal cycle (Figure 2; NASA, n.d). Peak salinity occurred at  
269 most sites mid-late year (mainland sites averaged  $227.1 \pm 27.0$  year day and ocean inlet and mid-  
270 lagoon sites averaged  $219.0 \pm 56.4$  year day) (Supplemental Table S2). The mainland sites had  
271 more variable maximum values (average of  $31.16 \pm 0.76$  psu) and statistically significant higher  
272 seasonal amplitudes (average of  $1.61 \pm 0.84$  psu,  $p=0.036$ ), likely related to their differences in  
273 freshwater input and/or longer residence times, while ocean inlet and mid-lagoon sites had

274 consistently higher maximum values (average of  $31.51 \pm 0.06$  psu) and lower seasonal  
275 amplitudes (average of  $0.45 \pm 0.06$  psu), due to ocean flushing (Supplemental Table S2).

### 276 *Dissolved Oxygen*

277 For dissolved oxygen, all sites were dominated by the first harmonic, with the dip in the  
278 summer concentrations due to the inverse relationship of solubility with temperature as well as  
279 annual factors such as air temperature, circulation, vertical mixing, air-sea gas exchange,  
280 photosynthetic oxygen production, and use of oxygen in decomposition (Figure 2; Kim et al.,  
281 2018). Additionally, benthic primary producers, such as seagrass meadows (Berg et al. 2019) and  
282 microalgae (McGlathery et al., 2001), can have important impacts on the fluctuations of water  
283 column oxygen concentrations that vary seasonally. Mainland sites generally had earlier to mid-  
284 year dates of minimum dissolved oxygen (average of  $216.4 \pm 8.6$  year day) and more variable  
285 minimum dissolved oxygen values (average of  $5.56 \pm 0.42$  mg/L) and seasonal amplitudes  
286 (average of  $2.56 \pm 0.12$  mg/L) due to longer residence times and warmer temperatures  
287 (Supplemental Table S3; Safak et al., 2018). Ocean inlet and mid-lagoon sites had higher  
288 minimum values (average of  $6.09 \pm 0.36$  mg/L) and more low to mid-range seasonal amplitudes  
289 (average of  $2.56 \pm 0.16$  mg/L) due to more consistently cool temperatures and higher flushing  
290 rates (Supplemental Table S3; Safak et al., 2018).

### 291 *Chlorophyll-a*

292 For  $\log_{10}(\text{Chl})$ , mainland sites generally exhibited a mix of annual and semi-annual  
293 seasonal harmonics, while ocean inlet and mid-lagoon sites were generally dominated by the first  
294 harmonic (Figure 2). Seasonal cycles of chlorophyll can be impacted by nutrient inputs,  
295 temperature, light availability (del Carmen Jiménez-Quiroz et al., 2021), as well as tide mixing,  
296 seasonal winds, upwelling, and stratification (Robles-Tamayo et al., 2020). In the VCR lagoons,

297 the concentrations seem to be most impacted by hydrology, such as addition of nutrients through  
298 groundwater, and atmospheric deposition (McGlathery et al., 2007), which due to the shallow  
299 nature of coastal lagoons, tended to stay for longer, and could cause lingering elevated algal  
300 concentrations (Gilbert et al., 2014; Tyler et al., 2003; Anderson et al., 2003, 2010; McGlathery  
301 et al., 2007). Sites exhibited geographic differences in the date for maximum, as mainland sites  
302 had an average of  $214.8 \pm 5.9$  year day and ocean inlet and mid-lagoon sites had a larger spread  
303 in dates with an average of  $234.2 \pm 29.4$  year day (Supplemental Table S4). The mainland sites  
304 had statistically significant higher maximum values (average of  $9.64 \pm 2.37$  ug/L,  $p=0.034$ ) and  
305 seasonal amplitudes (average of  $2.60 \pm 0.51$  ug/L,  $p=0.009$ ) likely related to longer residence  
306 times, allowing for more stagnant nutrient-rich waters (Supplemental Table S4; Safak et al.,  
307 2018). Ocean inlet and mid-lagoon sites had lower maximum values and seasonal amplitudes  
308 (averages of  $6.24 \pm 1.17$  ug/L and  $1.65 \pm 0.17$  ug/L) likely due to more mixing that dampens  
309 extremes (Supplemental Table S4; Safak et al., 2018).

### 310 *AOU*

311 For AOU, all sites were dominated by the first annual harmonic (Figure 2). AOU is  
312 related to many of the same factors as dissolved oxygen, as well as biological activity (BCO-  
313 DMO, 2023). During summer, lower solubilities from warmer water temperatures drive lower  
314 dissolved oxygen (Boyer et al., 1999). In previous studies, higher AOU values were found in the  
315 summer and fall months, with values closer to zero in the spring and winter when photosynthesis  
316 and respiration are low and roughly in balance and when air-sea gas exchange acts to resets AOU  
317 towards zero (Calleja et al., 2019). In general, mainland sites had later and more variable dates of  
318 maximum values (average of  $240.8 \pm 26.5$  year day), while ocean inlet and mid-lagoon sites  
319 were earlier (average of  $211.5 \pm 19.2$  year day) (Supplemental Table S5). The mainland sites had

320 mid to higher maximum values (average of  $1.13 \pm 0.32$  mg/L) and seasonal amplitudes (average  
321 of  $0.77 \pm 0.08$  mg/L), indicating higher summer net community respiration, potentially related to  
322 proximity to marsh and organic inputs to the lagoon that fuel bacterial respiration (Supplemental  
323 Table S5; Ducklow and Doney, 2013). Values at ocean inlet and mid-lagoon sites were generally  
324 slightly lower (averages of  $0.63 \pm 0.40$  mg/L and  $0.70 \pm 0.13$  mg/L) where sites are better  
325 flushed and away from marsh organic carbon inputs (Supplemental Table S5).

### 326 *Long Term Changes*

327         Statistically significant long-term trends were found for only a sub-set of the water  
328 quality variables at some VCR LTER sampling sites (Figure 3 and Supplemental Table S7).  
329 There were no distinct trend patterns either by water quality variable or by geography, with  
330 statistically significant trends occurring for salinity, AOU, and  $\log_{10}(\text{Chl})$  at both mainland and  
331 ocean inlet–mid-lagoon sites; the exception was temperature, which exhibited significant  
332 warming trends only at a cluster of mainland sites. Earth system modeling studies indicate the  
333 detection of ocean climate change trends requires 20-30 years for temperature and even longer  
334 (+50 years) for biogeochemical variables such as surface chlorophyll (Schlunegger et al., 2020).  
335 The absence of a statistically significant trend for many variables/sites (Supplemental Table S7)  
336 could reflect either a true lack of trend or detection issues because the signal to noise was small  
337 at sites sampled at low frequency with large seasonal and interannual variability (Henson et al.,  
338 2010); resolution of this issue will benefit in the future from longer coastal water quality time-  
339 series sampled at higher frequency.

340         Eight of the VCR LTER sites exhibited statistically significant, positive trends in  
341  $\log_{10}(\text{Chl})$  (Figure 3). Locations with significant positive  $\log_{10}(\text{Chl})$  trends were split evenly  
342 between mainland ocean inlet and mid-lagoon sites in the northern portion of the VCR LTER

343 sampling site, with no clear mainland-lagoon geographic pattern in trend magnitude  
344 (Supplemental Table S7). At the mainland northern VCR cluster of sites (RB, PCM, and  
345 RBCM), there was a co-occurrence of statistically significant, increasing temperature (mean  
346  $0.071 \pm 0.037$  °C/year ) and  $\log_{10}(\text{Chl})$  (mean  $0.014 \pm 0.007$   $\log_{10}(\text{ug/L})/\text{year}$  (errors on multi-  
347 site mean trends computed by propagating regression slope errors for each site assuming  
348 regression errors are independent),) consistent with warmer temperatures stimulating algal  
349 growth (Denchak, 2019).

350 Other findings from more nutrient enriched Maryland/Virginia coastal lagoons included  
351 Chl increasing in the lower part of the study area, close to the VCR region (Gilbert et al., 2014;  
352 Wazniak et al., 2007). Southern Mid-Atlantic coastal bodies of water had consistently high  
353 eutrophic conditions, as well as elevated levels of chlorophyll-a, with coastal lagoons being more  
354 impacted than river estuaries (Bricker, 2007). Chl concentrations could also increase in this area  
355 as macroalgal declines (McGlathery et al., 2001) or crashes, which can double the water column  
356 Chl concentrations (Tyler et al., 2001).

357 The changes in Chl were not limited to mainland sites, as  $\log_{10}(\text{Chl})$  also exhibited  
358 statistically significant, positive trends at four ocean inlet and mid-lagoon sites (mean  $0.017 \pm$   
359  $0.005$   $\log_{10}(\text{ug/L})/\text{year}$ ) (Supplemental Table S7). This could be related potentially to variations  
360 in water residence times at these sites due to changing hydrological factors (Denchak, 2019),  
361 though these is insufficient temporal information from simulated residence times, available for  
362 only two time periods 2002 and 2009, to indicate any long-term trends (Safak et al., 2015; Safak  
363 et al., 2018).

364 Five VCR LTER sites exhibited statistically significant, positive trends in salinity (Figure  
365 3). Salinity increased at three mainland sites (mean  $0.085 \pm 0.040$  psu/year) and at two ocean



366 inlet and mid-lagoon sites (mean  $0.068 \pm 0.026$  psu/year). These trends could be related to  
367 decreased freshwater input due to droughts or low streamflow, higher inundation of sea-level, or  
368 more intense storms causing breaching and wash over events that can vary by site (Supplemental  
369 Table S7; Anthony et al., 2009). However, the underlying driving factors are difficult to  
370 reconstruct because there are no USGS gauged streams for the VCR LTER region and only a  
371 single NOAA tide station (Wachapreague, VA, ID: 8631044), north of the VCR LTER sampling  
372 region, where local sea-level increased at  $5.63 \pm 0.59$  mm/year over the past three and a half  
373 decades (NOAA, 2024). Coastal storms have been found to have large impacts on salinity,  
374 especially when storm surges and waves are more intense for storms occurring during high tide  
375 (Kurylyk and Smith, 2023).

376         Statistically significant temperature increases were limited to three previously mentioned  
377 mid-VCR mainland sites (mean  $0.071 \pm 0.037$  °C/year), however many other sites had positive,  
378 though statistically insignificant trends (Supplemental Table S7). Previous studies have found  
379 regional surface ocean and air warming trends and increased marine heatwaves likely associated  
380 with anthropogenic climate change (Wiberg et al., 2023). Including all VCR LTER sites  
381 including sites with statistically insignificant trends (Table S7) resulted in a regional-mean  
382 warming trend of  $0.041 \pm 0.019$  °C/year that is broadly consistent with the measured trends  
383 (1982-2021) for nearby coastal ocean ( $0.030 \pm 0.016$  °C/year) and coastal bay ( $0.021 \pm 0.015$   
384 °C/year) weighted to the Wachapreague site (Wiberg et al., 2023). Coastal water warming trends  
385 have been found to have a positive correlation with regional atmospheric and oceanic  
386 temperatures on both monthly and decadal time scales (Najjar et al., 2010).

387         Statistically significant negative AOU trends were found at two sites, one mainland and  
388 one ocean inlet and mid-lagoon site ( $-0.057 \pm 0.054$  and  $-0.044 \pm 0.043$  mg/L/year, respectively)

389 (Supplemental Table S7). These AOU trends could be related to a change in lateral processes  
390 linking biogeochemical dynamics, organic carbon transport and freshwater flow (Figure 3).  
391 Murray et al. (2020) found that AOU is more influenced by mixing of end-member waters with  
392 different AOU than within-estuary biology, which could help to explain our findings.

393 Temporal shifts in seasonal harmonic elements were found when the long-term VCR-  
394 LTER time series was broken into earlier and later time periods (Supplemental Tables S9 and  
395 S10). For temperature, dates of maximums generally shifted earlier with increased maximum  
396 values and seasonal amplitudes; the shifts in seasonal amplitudes were significant statistically for  
397 mainland and all-sites. This aligns with shifting phenology due to global climate change  
398 (Anthony et al., 2009). For salinity, seasonal amplitudes increased with statistical significance at  
399 mainland, ocean inlet-lagoon, and all sites. This aligns with the shift of salinity due to changes in  
400 freshwater hydrology, evaporation and runoff, and groundwater, and threat of salt water intrusion  
401 from the nearby ocean (Anthony et al., 2009). For dissolved oxygen, dates of minimum values  
402 tended to shift earlier, while seasonal amplitudes increased, though neither of these signals was  
403 significant statistically for site types. The shift earlier of the dates of minimum value aligns with  
404 shifting earlier of maximum temperature values. For  $\log_{10}(\text{Chl})$ , statistically significant increases  
405 in maximum values were found for ocean inlet-lagoon and all sites. Seasonal amplitudes also  
406 decreased with statistical significance at several sites, which could be related to continuously  
407 elevated chlorophyll concentrations due to potentially more stagnant waters and longer residence  
408 times (Bricker, 2009). For AOU, dates of maximum values shifted earlier (significant  
409 statistically for ocean inlet-lagoon sites), while seasonal amplitudes increased. AOU shifting  
410 earlier is likely related to the shifting of increased biological effects (Ganguly et al., 2015).

411 *Simulated Low-Frequency Sampling*

412 Sub-sampling experiments were conducted to evaluate the ability of the low-frequency  
413 VCR-LTER sampling to capture the climatological annual cycle. When randomly subsampling  
414 the sites with high-frequency data sampling at the rate of low-frequency mainland  
415 measurements, the resulting harmonic fits tended to show relatively low average nRMSE values,  
416 means of  $0.692 \pm 0.306$  and  $0.620 \pm 0.220$  for Wachapreague and Willis Wharf, respectively  
417 (Supplemental Table S6). The relatively low nRMSE values indicated reasonable agreement  
418 between the low-frequency and high-frequency harmonic models.

419 Additionally, the low-frequency harmonics exhibited relatively good NSE values  
420 compared against the high-frequency harmonics, with averages of  $0.751 \pm 0.249$  and  $0.824 \pm$   
421  $0.172$  for Wachapreague and Willis Wharf respectively. The NSE values for all variables except  
422 for  $\log_{10}(\text{Chl})$  were within the acceptable range (above 0.65) at Wachapreague, with temperature  
423 ( $0.991 \pm 0.0005$ ) and dissolved oxygen ( $0.943 \pm 0.003$ ) being very good (Supplemental Table  
424 S6). At Willis Wharf, the NSE values for all variables besides AOU were within the acceptable  
425 range, with temperature ( $0.992 \pm 0.0004$ ), salinity ( $0.943 \pm 0.003$ ), and dissolved oxygen ( $0.940$   
426  $\pm 0.003$ ) all being well above the 0.65 threshold (Supplemental Table S6).  $\log_{10}(\text{Chl})$  had the  
427 lowest average NSE at Wachapreague ( $0.296 \pm 0.027$ ) and highest nRMSE ( $1.217 \pm 0.034$ ),  
428 while AOU had the lowest average NSE ( $0.532 \pm 0.027$ ) and highest nRMSE ( $0.965 \pm 0.025$ ) at  
429 Willis Wharf, with NSE values for both falling below the adequate threshold (0.65)  
430 (Supplemental Table S6). Both AOU and  $\log_{10}(\text{Chl})$  had slightly more complex seasonal cycles,  
431 with  $\log_{10}(\text{Chl})$  having a mix of annual and semi-annual harmonics, as discussed above in the  
432 *Chlorophyll-a* Results subsection and Figure 2. Therefore, AOU and  $\log_{10}(\text{Chl})$  may have been  
433 harder to capture more precisely at subsampled rates. Overall, for most of the water quality  
434 variables the nRMSE values were relatively low and the NSE values, for the most part, were in

435 acceptable ranges, indicating the skill of long-term VCR-LTER time-series in estimating  
436 seasonal cycles.

437         The second set of subsampling experiments with varying duration of quarterly sampling  
438 showed higher nRMSE values for 5-10 years of sampling, indicating that sites with minimal data  
439 availability would not resolve the climatological cycle well (Figure 4; Supplemental Tables S11  
440 and S12). After 50 years of quarterly sampling, Wachapreague had nRMSE across all variables  
441 of  $0.298 \pm 0.181$  and Willis Wharf had  $0.198 \pm 0.077$ . Across variables, sites, and harmonic  
442 elements, the duration of quarterly sampling to reduce nSD and reach the high-frequency  
443 confidence intervals varied from number of years to reach within the full harmonic confidence  
444 intervals varied from 10 to 23 years. It is important to note that for sub-sampling of  
445 Wachapreague  $\log_{10}(\text{Chl})$ , harmonic elements' confidence intervals were only crossed at one  
446 point for the date and value of maximum, and the seasonal amplitude. This could be because of  
447 the routine incorrect values at the site that may have been missed by the outlier removing  
448 processes. Overall, sub-sampling at Wachapreague indicated the need for slightly higher number  
449 of years of quarterly sampling for the sub-sampled low-frequency and high-frequency  
450 confidence intervals of harmonic elements ( $17.2 \pm 4.2$  and  $15.7 \pm 3.0$  years respectively)  
451 compared to Willis Wharf ( $14.9 \pm 4.6$  and  $17.2 \pm 3.3$  years respectively) (Supplemental Tables  
452 S11 and S12).

453         Shorter durations of quarterly sampling were required for certain variables to meet the  
454 high-frequency harmonic confidence intervals, with average value of  $13.3 \pm 3.6$  years for  
455 dissolved oxygen at Wachapreague and  $10.0 \pm 7.9$  years for temperature at Willis Wharf  
456 (Supplemental Tables S11 and S12). The longest duration of quarterly sampling to reach the  
457 high-frequency confidence intervals was  $23.3 \pm 11.3$  years for salinity at Wachapreague and

458 AOU at Willis Wharf, an average of  $23.0 \pm 6.3$  years (Supplemental Tables S11 and S12).  
459 Wachapreague's salinity values could be related to the fact that it had a more complex seasonal  
460 cycle, and therefore is harder to capture in fewer years. Willis Wharf's AOU had the highest  
461 nRMSE compared to the full harmonic curve in subsampling, indicating that it takes more data  
462 points to accurately complete its harmonic curve (Supplemental Table S6).

### 463 **Conclusions**

464 Using harmonic analysis, coastal water quality variables - temperature, salinity, dissolved  
465 oxygen,  $\log_{10}(\text{Chl})$ , and AOU - showed strong seasonality and geographic variation in the  
466 lagoon-barrier island system on the Eastern Shore of Virginia. The seasonal cycles of  
467 temperature, dissolved oxygen, and AOU in the VCR coastal lagoons were all dominated by an  
468 annual harmonic cycle, while salinity and  $\log_{10}(\text{Chl})$  had a mix of annual and semi-annual  
469 harmonic cycles. Mainland sites generally had higher maximum temperature,  $\log_{10}(\text{Chl})$ , and  
470 AOU values, lower dissolved oxygen values, and more complex salinity seasonal cycles than  
471 ocean-inlet and mid-lagoon sites due to longer residence times, shallower waters, and adjacency  
472 to marshes and uplands.

473 After removal of these seasonal cycles, linear regression analysis showed that all the  
474 VCR coastal lagoon water quality variables exhibited significant long-term trends, except for  
475 dissolved oxygen. Coastal water quality thus was not static to natural decadal variability and  
476 changing global climate conditions. Historical data analyses of multi-decadal time-series, such as  
477 for the VCR coastal lagoons here, complement simulation-based approaches for determining the  
478 sample density and duration required to detect climate change signals. By dividing the VCR  
479 LTER time-series into early and later time periods of roughly a decade each, changes in seasonal  
480 amplitude and phenology were identified from temporal shifts in some water quality harmonic

481 elements. These shifts generally involved increases in seasonal amplitudes for most variables and  
482 earlier of dates within the year for seasonal minimums and maximums. As illustrated here, the  
483 coupling of harmonic and regression analyses provides a compact and consistent statistical  
484 approach for characterizing seasonal variability, long-term trends, and shifting phenology at  
485 coastal monitoring and research sites more generally.

486         The seasonal harmonics were captured relatively well for all variables when sites  
487 measured at a high-frequency were subsampled at the equivalent seasonal resolution of the full  
488 VCR-LTER, low-frequency time-series. This indicated that the multi-decade VCR-LTER low-  
489 frequency water quality sampling approach generated consistent and sufficiently dense data sets  
490 to estimate the seasonal cycles using harmonic analysis. Based on sub-sampling of high-  
491 frequency time series, the average duration of quarterly sampling needed to reach the confidence  
492 intervals for the water quality variables ranged from 10-23 years, indicating that below 10 years  
493 of low-frequency sampling, sampling is unlikely to resolve the climatological seasonal cycle.  
494 The nRMSE of these variables all plateaued by 50 years, indicating that 50 years of low-  
495 frequency sampling results in robust and accurate harmonic fits to the climatological seasonal  
496 cycle.

497         The VCR-LTER project has long-term water quality data sets that range up to 30 years,  
498 within the estimated decade to multiple decade time window from the statistical sub-sampling of  
499 the high-frequency data. This suggests that the long data record from the VCR-LTER water-  
500 quality sampling scheme can be used to characterize well the seasonal variability across all  
501 variables investigated. A caveat is that long-term climate change trends may have already altered  
502 seasonal amplitudes and phenology, as suggested by statistically significant differences found for  
503 some harmonic parameters when the times-series was sub-set into earlier and later periods.

504           The VCR coastal data analysis presented here illustrated the value of even a few years of  
505 high-frequency water-quality sonde data sets as a complement to more traditional and widely  
506 used manual data collection approaches. The information gained from coupling harmonic  
507 analysis with statistical sub-sampling of high-frequency records can guide researchers analyzing  
508 existing and historical time-series and establishing new water-quality monitoring sites. The  
509 statistical sub-sampling analysis highlighted clearly the trade-offs of sampling frequency versus  
510 duration or sampling density for identifying seasonal variation in the VCR coastal lagoons. The  
511 same approach can be applied more generally to other coastal sites and to other water quality  
512 variables that are not yet measured or just beginning to be measured at long term sites sampled at  
513 a low-frequency, using a reference station measured at a high-frequency nearby if available.

514

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### 519 **Author Contribution Statement**

520           E. I. B. initially devised the methodology, wrote the code and performed all of the  
521 analysis, and led the writing of the article. S. C. D. and K. J. M. contributed to method  
522 development, data analysis, and writing and editing of the article.

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527 **Conflict of Interest Statement**

528 The authors declare no conflict of interest.

529 **Data Availability Statement**

530 High-frequency water quality data were provided by the College of William and Mary's  
531 Virginia Institute of Marine Science Eastern Shore Laboratory (VIMS ESL) with the assistance  
532 of ESL's Darian Kelley. Data can be retrieved from the VIMS ESL website  
533 ([https://www.vims.edu/esl/research/water\\_quality](https://www.vims.edu/esl/research/water_quality)).

534 Low-frequency water quality data (doi: knb-lter-vcr.247.17) were retrieved from the  
535 VCR website ([https://www.vcr.lter.virginia.edu/cgi-bin/showDataset.cgi?docid=knb-lter-](https://www.vcr.lter.virginia.edu/cgi-bin/showDataset.cgi?docid=knb-lter-vcr.247)  
536 [vcr.247](https://www.vcr.lter.virginia.edu/cgi-bin/showDataset.cgi?docid=knb-lter-vcr.247)).

537

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739 **Tables and Figures**

Table 1. Names, site type, coordinates, and dates measured for each coastal water quality site in the Virginia Coast Reserve, where bolded sites are high-frequency ESL locations, and non-bolded sites low-frequency VCR-LTER locations.

Site	Site Type	Latitude (Degree)	Longitude (Degree)	Dates Measured
<b>Wachapreague (W)</b>	Mainland	37.608	-75.686	March 25th, 2016 to December 31st, 2022
<b>Willis Wharf (WW)</b>	Mainland	37.512	-75.806	October 12th, 2018 to December 24th, 2022
Ramshorn Channel Creek (RCC)	Mainland	37.303	-75.905	August 26th, 2004 to October 11th, 2022
Redbank Creek Mouth (RBCM)	Mainland	37.460	-75.816	August 26th, 2004 to October 10th, 2022
Cattleshed Creek Mouth (CCM)	Ocean Inlet	37.443	-75.689	July 28th, 1992 to October 10th, 2022
Little Cobb Island (LCI)	Ocean Inlet	37.305	-75.792	August 26th, 2004 to October 11th, 2022
Machipongo Inlet (MI)	Ocean Inlet	37.368	-75.736	July 31st, 1997 to October 10th, 2022
New Marsh (NM)	Mid-Lagoon	37.291	-75.857	August 26th, 2004 to October 10th, 2022
Oyster Harbor (OH)	Mainland	37.289	-75.924	July 28th, 1992 to October 11th, 2022
Phillips Creek Mouth (PCM)	Ocean Inlet	37.445	-75.834	July 28th, 1992 to October 10th, 2022
Quinby Inlet (QI)	Ocean Inlet	37.467	-75.668	August 25th, 2004 to October 10th, 2022

Red Banks (RB)	Mainland	37.464	-75.807	July 28th, 1992 to October 10th, 2022
South Hog (SH)	Ocean Inlet	37.382	-75.718	July 31st, 1997 to October 10th, 2022
Shoal Site (SHS)	Mid-Lagoon	37.417	-75.761	August 26th, 2004 October 10th, 2022
Sand Shoal Inlet (SS)	Ocean Inlet	37.290	-75.785	August 26th, 2004 to October 11th, 2022

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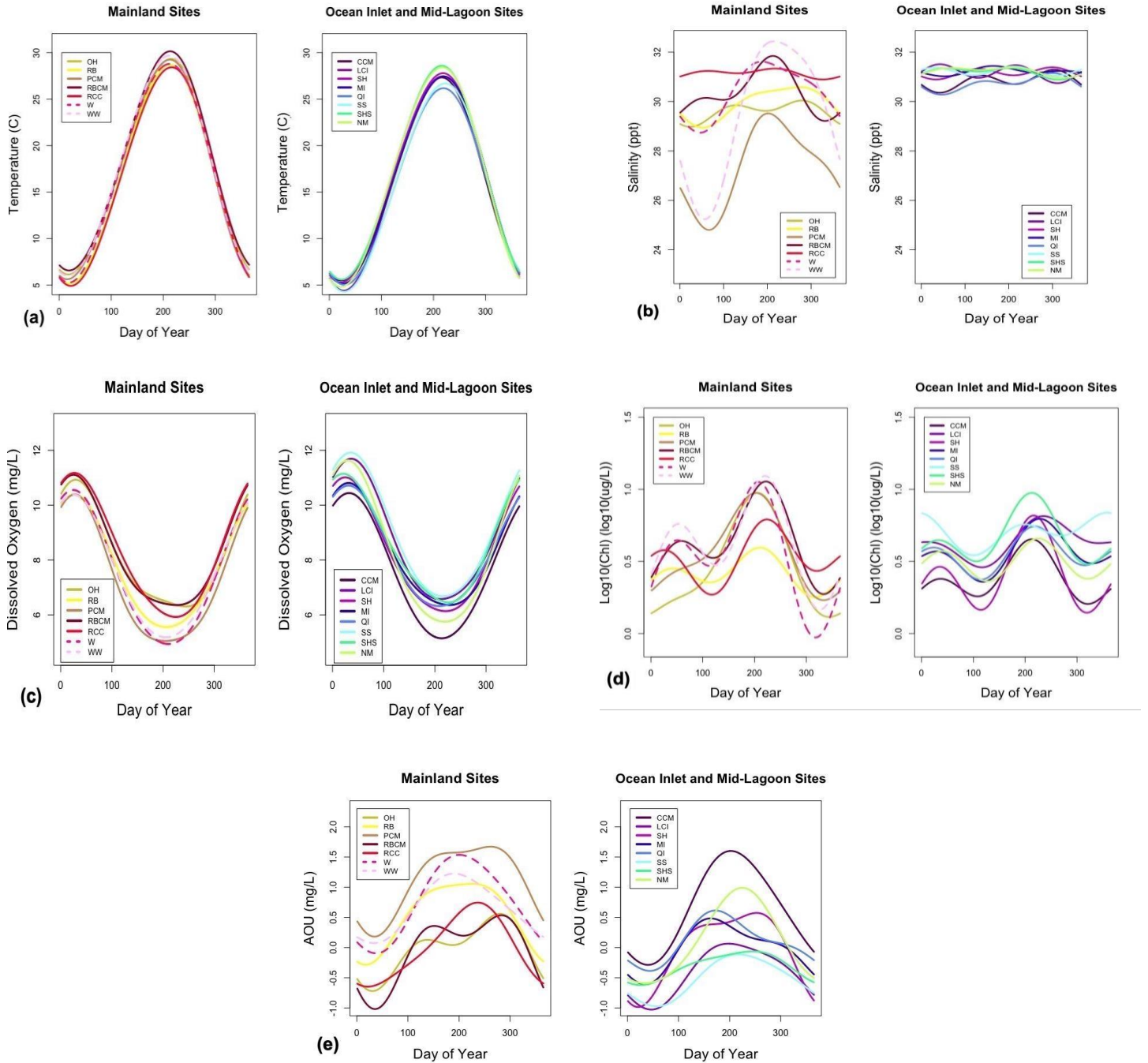
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744 **Figure 1.** Spatial map of the eastern shore of Virginia, United States created in ArcGIS using  
 745 Imagery (WGS84) base map showing the locations of the ESL sites as purple dots, and the VCR-  
 746 LTER sites as pink dots (see Table 1). The sites in the orange box are considered mainland, and  
 747 the sites in the teal box are considered ocean inlet and mid-lagoon sites (Table 1). The Virginia  
 748 Coast Reserve shallow lagoon-barrier island system is bounded to the west by the Eastern Shore  
 749 peninsula and to the east by barrier islands. The lagoon system is flushed by tidal flows from the  
 750 coastal Atlantic Ocean (right side of image) via ocean inlets between the barrier islands.

751



752 **Figure 2.** Temperature Composite harmonic fits for mainland sites (left panel) and ocean inlet  
 753 and mid-lagoon sites (right panel for). (a) Temperature (b) Salinity (c) Dissolved Oxygen (d)  
 754 Log<sub>10</sub>(Chl) and (e) AOU.

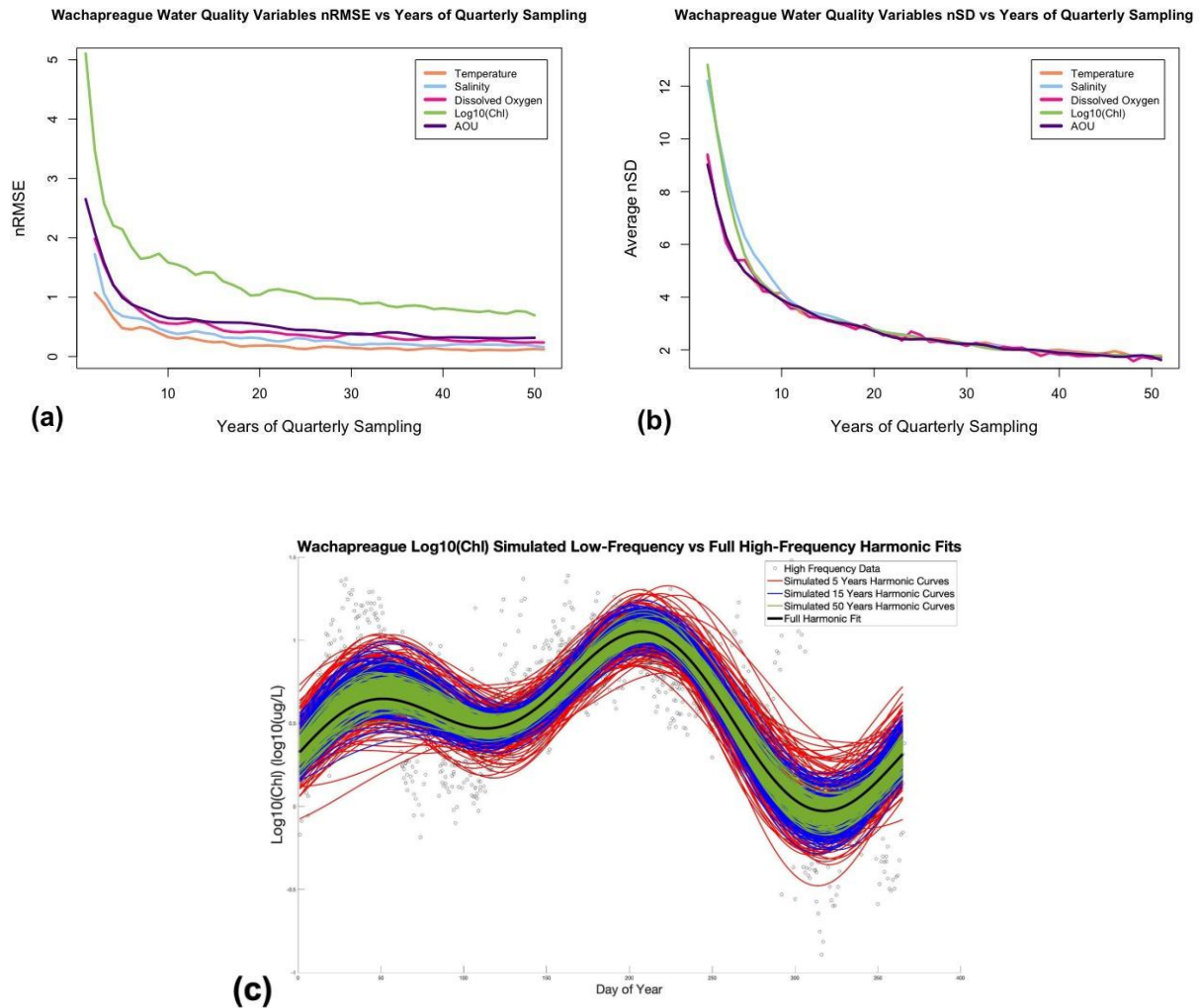
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757 **Figure 3.** Spatial map of statistically significant multi-year temporal trends in water quality  
758 variables at VCR-LTER and ESL sites.

759



760 **Figure 4.** Influence of quarterly sampling duration on low-frequency sub-sampling harmonic  
 761 fits versus high-frequency fits for (a) nRMSE, (b) average nSD of date of maximum value and  
 762 seasonal amplitude versus successive years of quarterly sampling at site Wachapreague, and (c)  
 763 harmonic curves for 200 trials of simulated low-frequency sampling of log10(Chl) at site  
 764 Wachapreague for sampling densities of 5 years (in red), 15 years (in blue), and 50 years (in  
 765 green) versus the full high-frequency harmonic fit (in black) (data points marked as circles).